

MANAGEMENT OF SEAWATER INTRUSION IN COASTAL AQUIFERS: PRESENT AND FUTURE CHALLENGES

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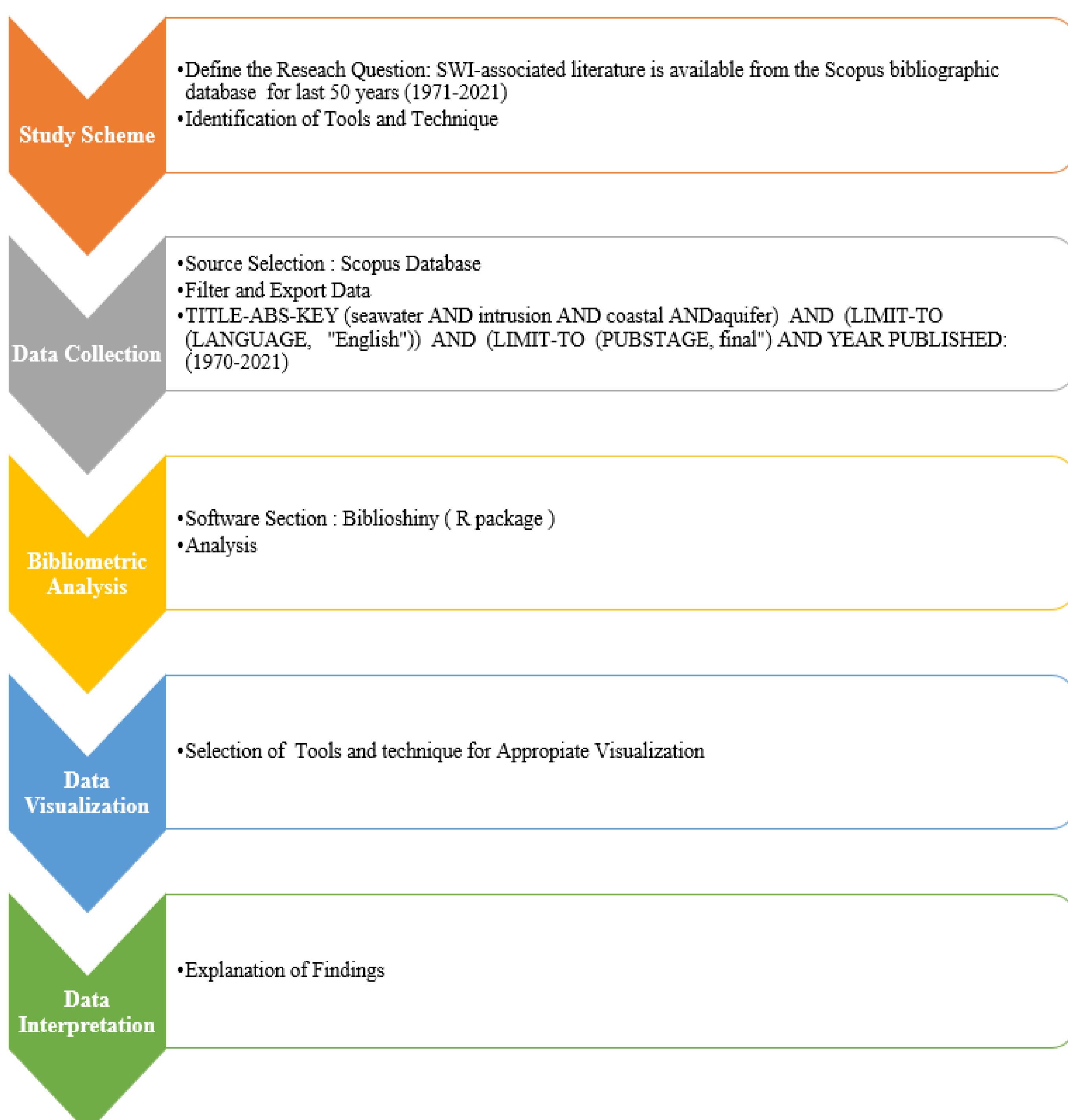
Background:

- Groundwater is considered the main source of water in many coastal areas. The increase of water demands increases the abstraction from aquifers which has resulted in lowering water tables and caused saltwater intrusion (SWI).
- Saltwater intrusion is one of the main causes of groundwater quality degradation and a major challenge in the management of groundwater resources in coastal regions.
- Understanding the current SWI situation and forming research consensus are of great significance for further studies and coastal water resources management.

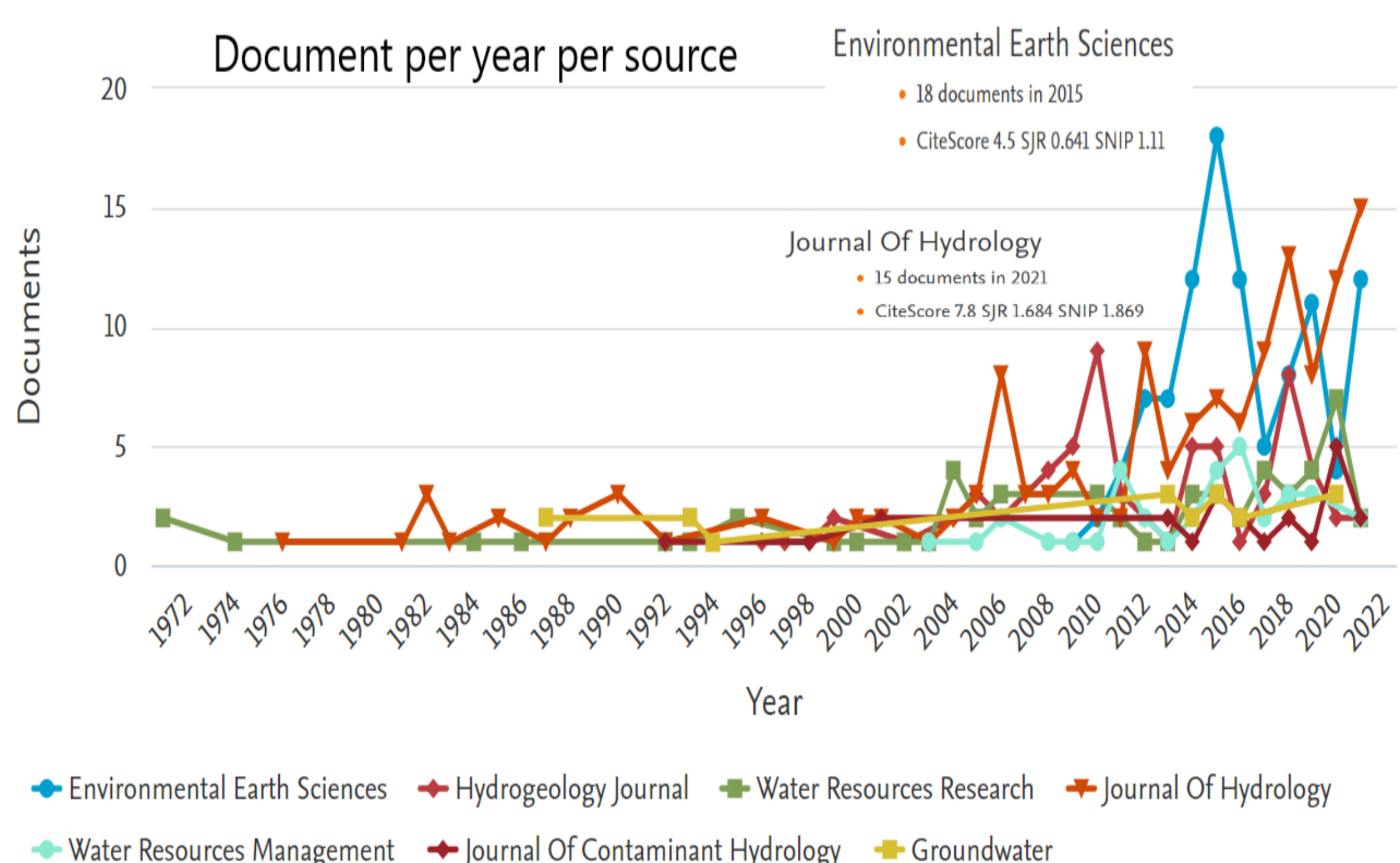
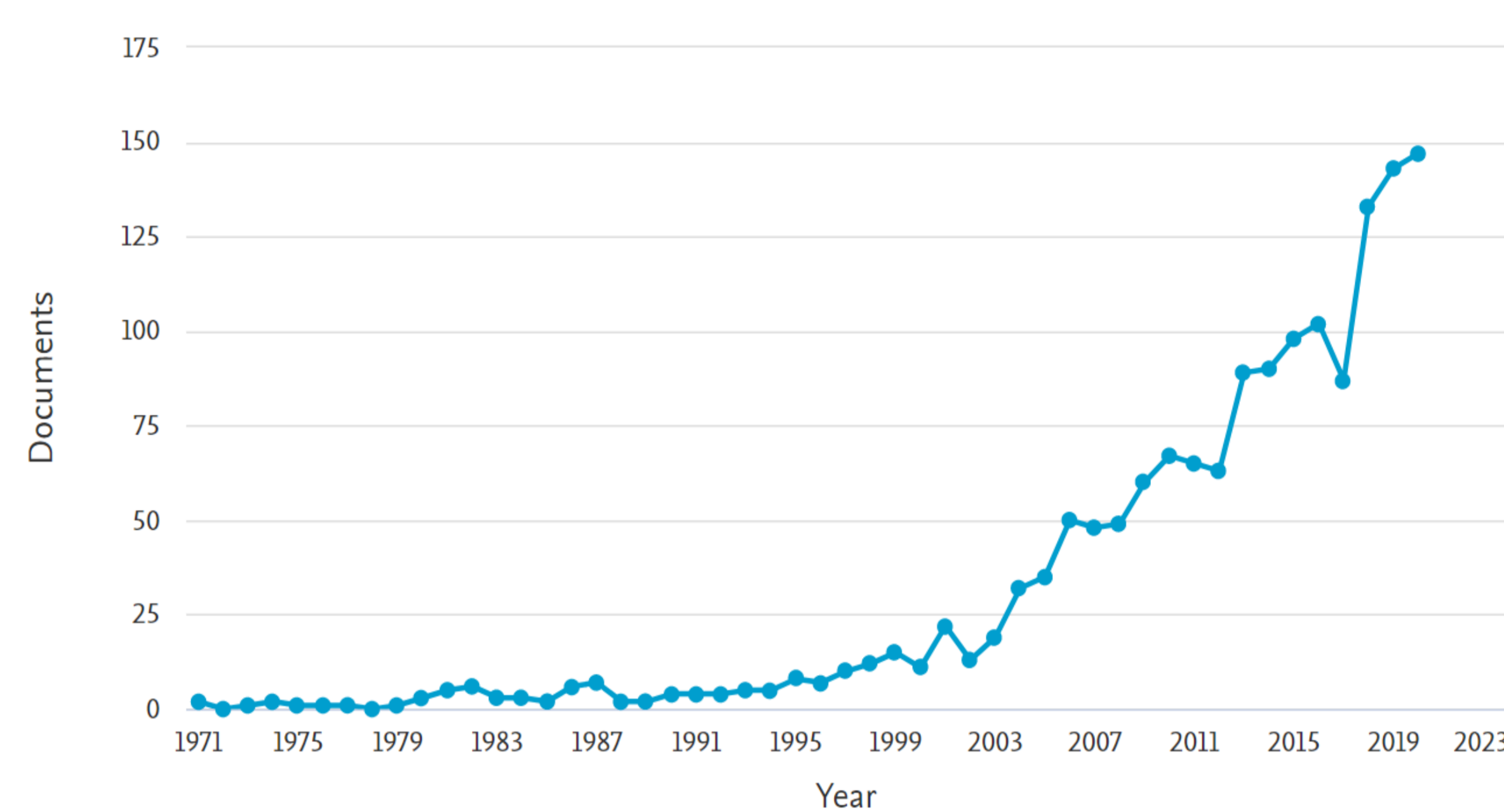
Objectives :

- This study conducted a bibliometric analysis to evaluate the SWI-associated literature is available from the Scopus bibliographic database for last 50 years (1970-2020).
- Statistical and scientific mapping analysis by using Biblioshiny excavate existing research achievements further.

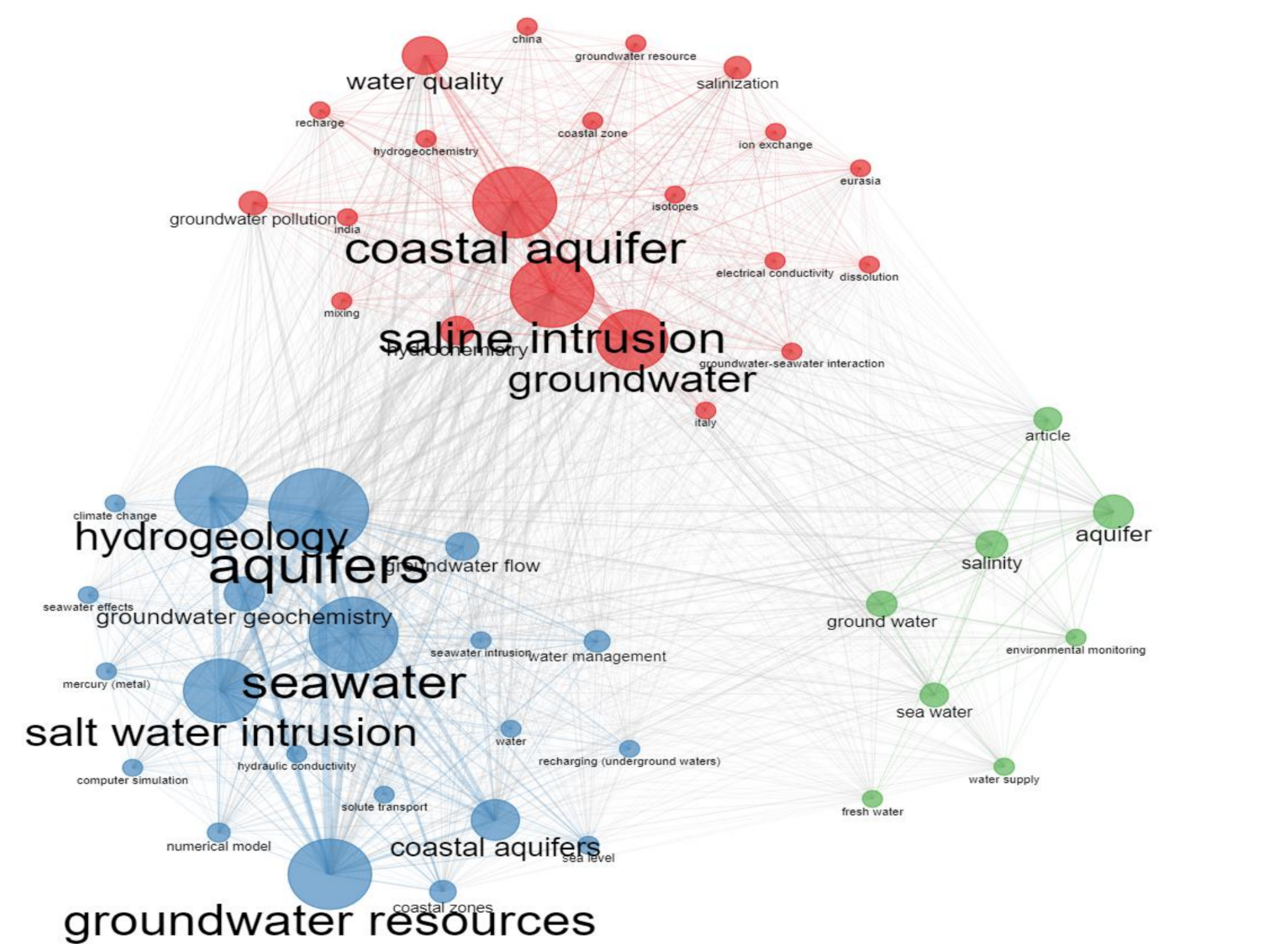
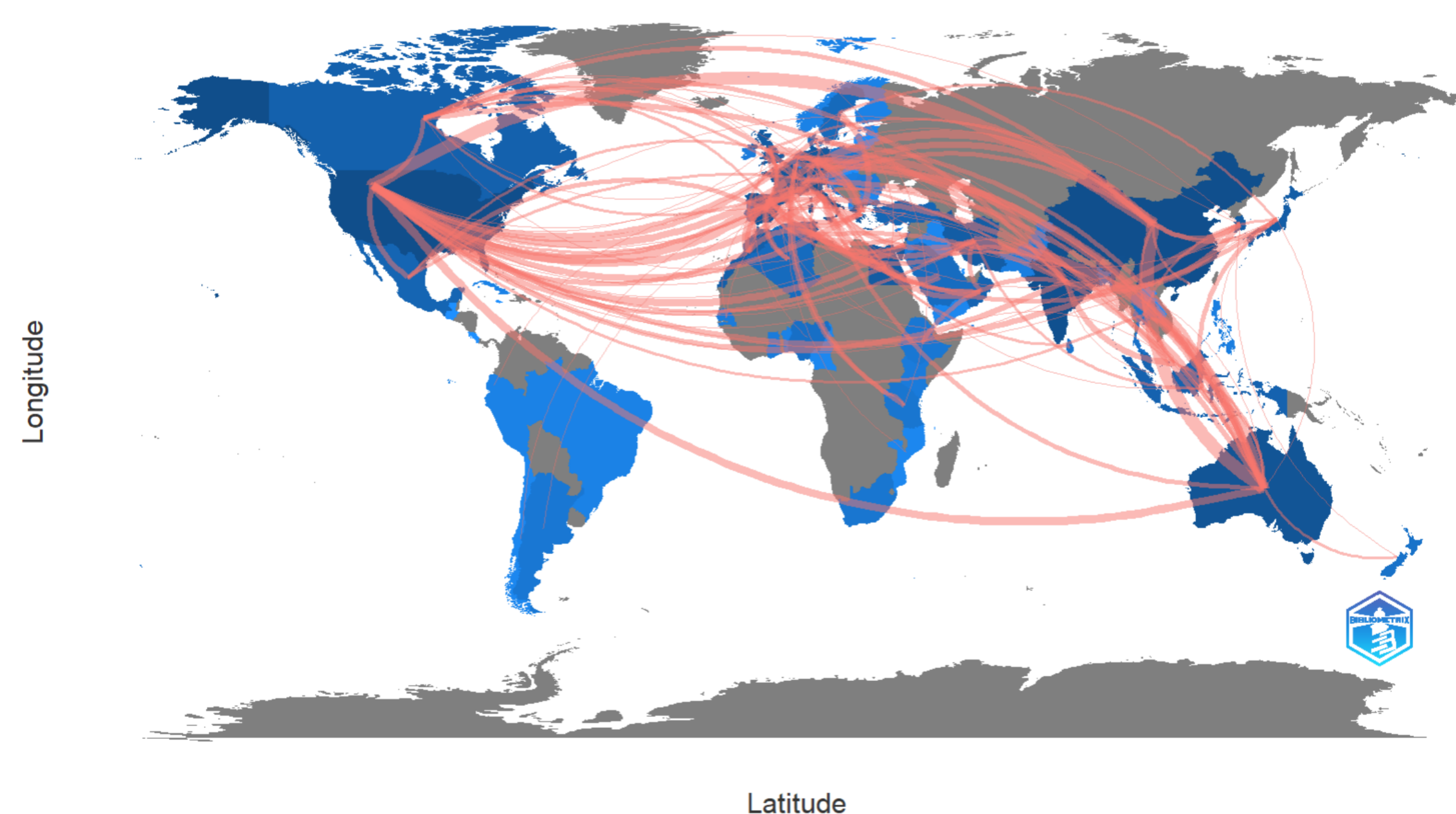
Standard Bibliometric Analysis Process



Documents by year



Country Collaboration Map



Observation:

- From 1971 to 2021, RTI research productivity gradually grew after that initial publication.
- More document was published in Environmental Earth Science in that Journal of Hydrology, but Citation score was more Journal Of hydrology than others.
- Therefore, the greater the CiteScore, the more critical the journal is considered. depict the most common keywords listed by authors in SWI research publications originating in the WORLD MAP.
- The circle size and colors denote keyword occurrence levels and different coupling clusters, respectively; keywords listed four or more times were included; 33 keywords met this criterion.
- The most common keywords in seawater intrusion of coastal aquifer publications were seawater (1067 occurrences), aquifer (1027 occurrences), and coastal aquifer (939 occurrences). Terms such as saline, climate change, etc., are least expected.

Conclusion:

- Implementing SWI knowledge, such as processes, investigative methodologies, and managerial approaches is perhaps the most critical scientific problem in SWI research. Strategies are required to close the gaps between SWI knowledge and management practice.
- To illustrate the benefits of SWI research methodologies and knowledge, more studies that retrospectively and critically analyze the effectiveness of management practices, particularly initiatives deriving from SWI research findings, are needed to illustrate the benefits of SWI research methodologies and knowledge.
- Subsurface intakes were not considered a mitigating strategy for seawater intrusion into costal aquifers in the available literature and experiments.

Management of Seawater Intrusion in Coastal Aquifers: Present and Future Challenges¹⁵⁶

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Keywords Seawater Intrusion, Coastal aquifer, Groundwater, Management, Bibliometric

Abstract

Due to population growth and rapid industrialization, massive withdrawals of groundwater resources have resulted in seawater intrusion into coastal aquifers worldwide. This problem is common in arid and semi-arid regions (40% of the world's population.) The effects of seawater intrusion on the local community's health and economic and socio-cultural developments in coastal areas have prompted a wide range of research. The current study's goal is to provide a comprehensive review of the processes that control seawater intrusion into coastal aquifers over the last 50 years and the necessary mitigation measures. The SWI-associated literature is available from the Scopus bibliographic database and statistical and scientific mapping analysis by using Biblioshiny to excavate existing research achievements further. This intends to examine the reputation and trends of seawater intrusion in coastal aquifer research from 1970 to 2021 to assist other researchers in comprehending the global scope of study and predicting the dynamic directions of seawater intrusion in coastal aquifer research. Implementing SWI knowledge, such as processes, investigative methodologies, and managerial approaches is perhaps the most critical scientific problem in SWI research. Strategies are required to close the gaps between SWI knowledge and management practice. To illustrate the benefits of SWI research methodologies and knowledge, more studies that retrospectively and critically analyze the effectiveness of management practices, particularly initiatives deriving from SWI research findings, are needed to illustrate the benefits of SWI research methodologies and knowledge. The following are the goal: a) Examine the patterns of seawater intrusion in journal publications about coastal aquifers, and b) Comprehensive review of the processes that control seawater intrusion into coastal aquifers and the new proposed methodology are then explained in detail.

1 **MANAGEMENT OF SEAWATER INTRUSION IN COASTAL AQUIFERS:**
2 **PRESENT AND FUTURE CHALLENGES**

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9 **KEYWORDS**

10 SEAWATER INTRUSION, COASTAL AQUIFER, HYDROGEOLOGY,
11 GROUNDWATER, SALINE, WATER MANAGEMENT

12 **ABSTRACT**

13 Due to population growth and rapid industrialization, massive withdrawals of groundwater
14 resources have resulted in seawater intrusion into coastal aquifers worldwide. This problem is
15 common in arid and semi-arid regions (40% of the world's population.) The effects of seawater
16 intrusion on the health of the local community and economic and socio-cultural developments
17 in coastal areas have prompted a wide range of research. The current study's goal is to provide
18 a comprehensive review of the processes that control seawater intrusion into coastal aquifers
19 over the last 50 years, as well as the necessary mitigation measures. Various investigation
20 methods and the new proposed methodology are then explained in detail.

21 **1 INTRODUCTION**

22 For domestic, industrial, and agricultural needs, most of these coastal regions rely on
23 groundwater as their primary source of freshwater. As the population continues to rise at an
24 alarming rate, increased demand for freshwater supplies is met by extensive pumping of fresh
25 groundwater, resulting in a lowering of the water table or piezometric head, which leads to
26 saltwater intrusion (SWI) and emphasizes the importance of groundwater monitoring,
27 management, and conservation(Bear, 1999). Around the world, more than 100 countries and
28 regions around the world are threatened by seawater intrusions such as the Nile Delta, the
29 Mekong Delta, East, and South Asia, the Mediterranean region, the Wadi Ham delta, and North
30 America(H. Abd-Elhamid et al., 2016; Barlow & Reichard, 2010; Han & Currell, 2018; M.
31 Sherif & Kacimov, 2007; Vu et al., 2018).

32 SWI is the subterranean migration of seawater into fresh aquifers, primarily driven by
33 unsustainable groundwater exploitation and linked to rising sea levels and climate change
34 (Werner et al., 2013). The significant increase in total dissolved solids (TDS) induced by
35 enrichment solutes in seawater, such as chloride or sodium, distinguishes SWI(Seenipandi et
36 al., 2019). The famous Ghyben-Herzberg formula was proposed over a century ago to
37 statistically examine the interaction between brackish and fresh groundwater in coastal zones
38 (Alexander Herzberg, 1901; Drabbe & Ghijben, 1889; Todorovic & Verruijt, 1968). It was
39 commonly employed to locate the fresh-saline water interface, governed by the unconfined

40 groundwater level. The limitations of this formula were then described mathematically by (Bear
41 1985; Hubbert, 1940). Researchers have been attempting to quantify the form of the saltwater
42 wedge and measure the degree of groundwater salinization since then (Post, 2018). (Bear, 1999)
43 addressed the variable density and hydrodynamic dispersion problem as a significant symbol
44 of SWI research. Coastal water management strategy, groundwater resource protection,
45 submarine groundwater discharge (SGD), field monitoring methods, hydrochemical and
46 isotopic concerns, and numerical simulation have evolved into a complex SWI research system.
47 The SWI process is becoming more widely known, particularly the interplay between surface
48 and groundwater, and seawater induction is now thought to significantly affect groundwater
49 salinization(Cao et al., 2020; Ferguson & Gleeson, 2012). Field study, laboratory
50 experimentation, analytical solution, and numerical simulations are some research approaches.
51 SWI researchers gathered massive data on a global scale and applied several innovative
52 technologies after years of geological survey and environmental monitoring. Finding hot
53 subjects and distinguishing potential breakthrough directions has become a worry for
54 scholars(Javadi et al., 2015; Mabrouk et al., 2018; Post, 2005).

55

56 Many types of software have been developed in recent years, allowing for increased
57 bibliometric use in various research areas (Garfield, 2009; Leydesdorff & Bihui, 2005). The
58 bibliometric method has been used in the field of ecology and hydrology to investigate advances
59 in a variety of topics, including global groundwater research (Niu et al., 2014), stable isotope
60 of precipitation research (Wei et al., 2019), urban development, and ecological safety research,
61 and global research in sustainable development (Qin, 2015). Scopus is the only database that
62 includes a comprehensive, curated abstract and citation database and enriched data and linked
63 scholarly content (Hassan et al., 2014). Find relevant and trustworthy research quickly, identify
64 experts, and get reliable data, metrics, and analytical tools to support confident research plan
65 decisions - all from a single database and subscription.

66 The bibliometric R-package software was used in this work, open-source software that provides
67 a set of tools for undertaking quantitative research in bibliometrics. Biblioshiny created the R-
68 package, written in the R programming language(Aria & Cuccurullo, 2017). It contains the
69 most important algorithms for statistical and scientific mapping analysis. The web interface app
70 (Biblioshiny) was introduced in recent versions of the bibliometrix R-package (i.e., 2.0
71 upwards) to assist users without coding abilities in conducting bibliometric analysis. Data can
72 be imported in BibTex, CSV, or Plain Text format from databases via the Biblioshiny interface.
73 Biblioshiny also allows you to filter data. Our research took advantage of the features of
74 biblioshiny for bibliometrix to import data from Scopus in BibTex format. In the results section,
75 the study analysis is presented. This review work intends to examine the reputation and trends
76 of seawater intrusion in coastal aquifer research from 1970 to 2021 to assist other researchers
77 in comprehending the global scope of research and predicting the dynamic directions of
78 seawater intrusion in coastal aquifer research. Implementing SI knowledge, such as processes,
79 investigative methodologies, and managerial approaches, into practice is perhaps the most
80 critical scientific problem in SI research. Strategies are required to close the significant gaps
81 between SI knowledge and management practice. To illustrate the benefits of SI research
82 methodologies and knowledge, more studies that retrospectively and critically analyze the
83 effectiveness of management practices, particularly initiatives deriving from SI research

84 findings, are needed to illustrate the benefits of SI research methodologies and knowledge. The
85 following are the goal:

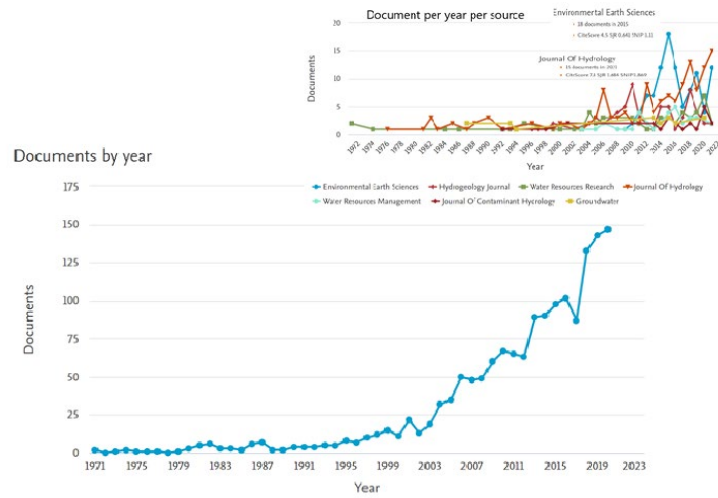
- 86 i. Examine the patterns of seawater intrusion in journal publications about coastal
87 aquifers.
- 88 ii. Comprehensive review of the processes that control seawater intrusion into coastal
89 aquifers and its advantages and disadvantages.
- 90 iii. New proposed method to control SWI.

91 **2. BIBLIOMETRIC ANALYSIS**

92 A bibliometric analysis study differs from a review paper in that it focuses on current trends,
93 difficulties, and future directions for a specific subject. Scopus' built-in tool (for extracting the
94 database) and Biblioshiny (mapping and visualization) software were used for this
95 bibliometric investigation. Tracing publications using the SCOPUS database is a systematic
96 approach to bibliometric study. The issue was searched in SCOPUS utilizing the search phrase
97 "Seawater Intrusion Coastal Aquifer." Only articles or ongoing papers were included in the
98 final results. The SCOPUS database yielded 1,683 accurate document records, according to the
99 findings. The year of publication is limited in this study, and it ranges from the earliest, in 1970,
100 to the most recent, in 2021. TITLE-ABS-KEY (seawater AND intrusion AND coastal AND
101 aquifer) AND (LIMIT-TO (LANGUAGE, "English")) AND (LIMIT-TO (PUBSTAGE,
102 "final")) AND YEAR PUBLISHED: (1970-2021) were the search parameters. As a result, 1,683
103 documents were shown. In addition, the majority of the document kinds were articles. Later
104 these documents are exported to Bibtex file format. The findings were reviewed based on the
105 year, source, author, affiliation, nation, study area, document types, languages, and the critical
106 topic's search. Bibliometric indicators like total publications, top institutions, leading authors,
107 and h-index were utilized to rank the papers. The data was analyzed using Biblioshiny, a
108 combination of spreadsheet and bibliometric analysis tools (R Package). Biblioshiny was used
109 to perform the author keyword analysis and bibliographic coupling. Biblioshiny was used to
110 generate global collaboration, subject trends, three-component analysis, word clouds, and
111 treemap analysis. The first publication on seawater intrusion in coastal aquifer research was in
112 1970, and it was mentioned 19 times.

113 From 1971 to 2021, RTI research productivity gradually grew after that initial publication.
114 Figure 1 shows the articles and cumulative publications records from 1970 to 2021, indicating
115 a substantial increase in overall accumulated publications. If this is the case, the annual
116 publication will continue to rise. Figure 1 represents several journal Documents per source
117 Journal. There is gradual flocculation occurring between the Environmental Earth Science and
118 Journal of Hydrology from 1971 to 2021. More document was published in Environmental
119 Earth Science in that Journal of Hydrology, but Citation score was more Journal Of hydrology
120 than others. Therefore, the greater the CiteScore, the more critical the journal is considered.

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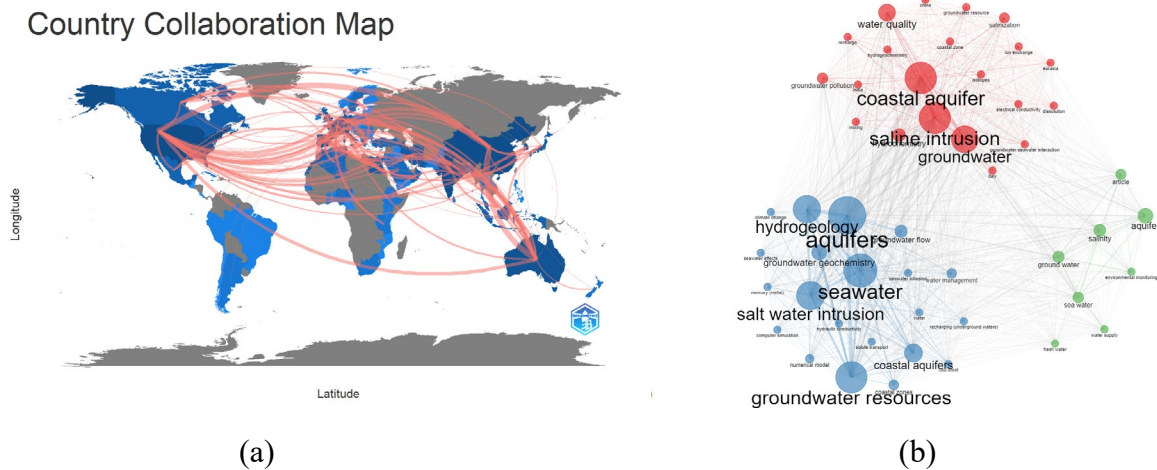


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Figure 1: Number of Journal Publication from 1971–2021 and Document Number per year per source.



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(a)

(b)

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Figure 2: Country Collaboration Map and common keywords listed in SWI research

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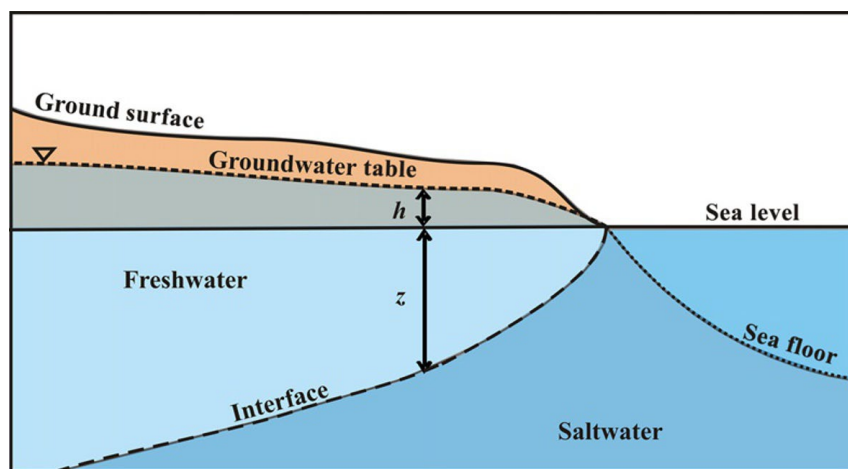
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Figure 2 (b) depict the most common keywords listed by authors in SWI research publications originating in the WORLD MAP. The circle size and colors denote keyword occurrence levels and different coupling clusters, respectively; keywords listed four or more times were included; 33 keywords met this criterion. The most common keywords in seawater intrusion of coastal aquifer publications were seawater (1067 occurrences), aquifer (1027 occurrences), and coastal aquifer (939 occurrences). Terms such as saline, climate change, etc., are least expected.

141 **3. SALTWATER AND FRESHWATER INTERACTION MECHANISM**

142 Saltwater intrusion is one of the world's most significant environmental problems, threatening
143 the quality and availability of freshwater in coastal aquifers, especially in arid and semi-arid
144 regions. In natural conditions, SWI refers to replacing freshwater in coastal aquifers with
145 seawater due to density-dependent landward migration of a salty water body SWI is defined by
146 (M. Sherif & Singh, 2002) as a mechanism in which dense seawater displaces deep inland
147 freshwater and moves along the bed's floor. As a result, it is regarded as the final result of
148 density-dependent interaction. Between freshwater and saltwater, it is responsible for
149 groundwater movement dynamic equilibrium. Although the hydraulic gradient of flow into the
150 sea is relatively small, the slight density gradient between seawater (density of 1025 kg/m³)
151 and freshwater (density of 1000 kg/m³) plays a significant role in the natural progression SWI.
152 The displaced freshwater with lower density shifts to upper regions and leaves the aquifer as
153 submarine groundwater discharges in a cyclic pattern to maintain dynamic equilibrium in the
154 system. As a result, the flow pattern in the system along the coast is cyclic (even under steady-
155 state conditions), with some saline water entrained within the overlying freshwater and returned
156 to the sea, causing further seawater intrusion(Barlow, 2003; M. Sherif & Singh, 2002).

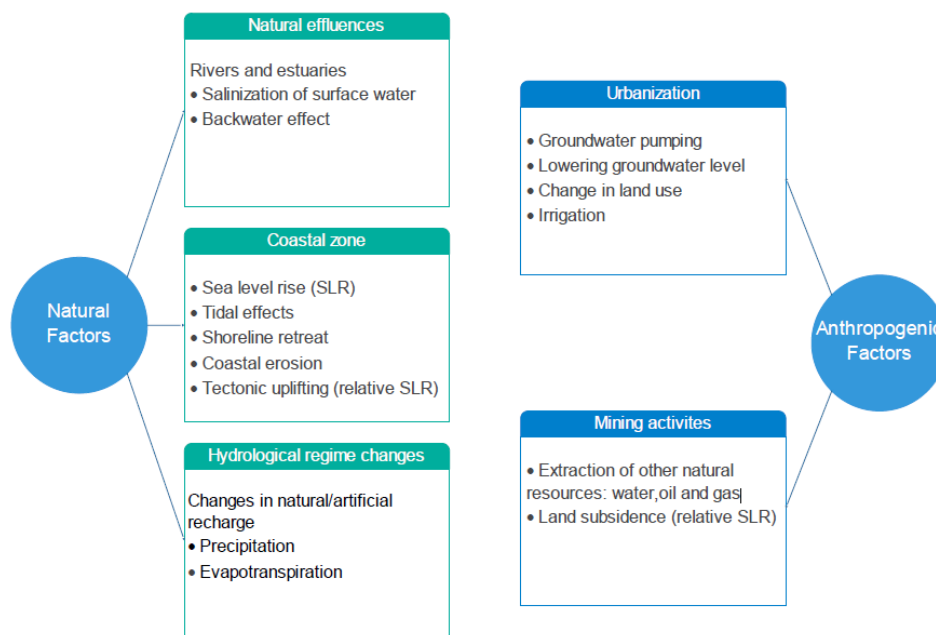
157 Furthermore, the spreading of the solute and the mixing phase is regulated by hydrodynamic
158 dispersion, combining mechanical and Physico-chemical distribution (molecular diffusion).
159 This process creates a mixing zone (or transition zone) between the two fluids, where the water
160 density changes from freshwater to saltwater (Bear, J. & Zhou, 2007; Bear, J. J. & Cheng, 1999;
161 Sherif, M. M. & Al-Rashed, 2001)In coastal systems, this region still occurs between freshwater
162 and salty water. It is thought to be the main flow route for the available cyclic movement of
163 saltwater receding to the sea (Bear, 1985). The saltwater/freshwater interface is the boundary
164 between the bodies of these two fluids. They formed a distinct zone of salty water in the
165 freshwater body is known as the regional saltwater wedge. As a result, the saltwater wedge is
166 designed to be a source of pollution that degrades aquifer water quality and quantity (Figure 3).
167 As shown in, any coastal aquifer with this issue should have three distinct zones: saltwater zone,
168 mixing/transition zone, and freshwater zone.



170 Figure 3: Saltwater and Freshwater Interaction(Vandenbohede et al., 2009)

171 **4. FACTORS INFLUENCING SALTWATER INTRUSION**

172 The common factor influencing saltwater intrusion is natural and anthropogenic factors. The
173 most common factors that are worsening the Saltwater intrusion (SWI) issue by disrupting the
174 natural hydraulic equilibrium condition in coastal aquifer systems due to natural and human
175 activities are summarized by (Oude Essink, 2001; Patel & Shah, 2008) in Figure 7. The
176 overall effect of these factors is to increase total aquifer salinization, which has significant
177 implications for the quality of groundwater supplies and the natural environment. Some of
178 these variables have short-term effects (for example, tidal), others have intermittent effects
179 (for example, seasonal variations in natural groundwater flow), and others have long-term
180 consequences (for example, climate change and the majority of human activities).



182 Figure 4: Natural and anthropogenic factors influencing the saltwater intrusion (Hussain,
183 2015)
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185 **5. METHOD USED FOR CONTROLLING THE SALTWATER INTRUSION**

186 Groundwater is a vital resource for supporting the societies and economies of coastal regions,
187 which are among the world's most heavily populated areas. One-third of the world's freshwater
188 demand is met by groundwater (Bear, 1999). (Howard, 1987) addressed the degrading effects
189 of SWI caused by excessive groundwater abstraction. The initial loss of freshwater can be offset
190 by natural replenishment of storage at regulated groundwater withdrawal rates. However, the
191 problem (SWI) only occurs when there is an inappropriate and unplanned use of fresh
192 groundwater combined with a substantial decline in recharge, resulting in public health,
193 socioeconomic, and environmental issues. As a result, remediation steps must be implemented
194 to prevent further deterioration in water quality in coastal aquifers. However, groundwater
195 remediation may be very expensive and time-consuming, depending on the source and degree
196 of salinity. As a result, various arrangements and prices of external sources/sinks of water will
197 promote sustainable water resource management in coastal areas. Determining a suitable

198 ultimate landward extent of saline water and the calculation of the sufficient freshwater
199 discharge needed to preserve the interface of the saltwater wedge in the seacoast location are
200 both parts of the management of coastal aquifers.

201 A seaward hydraulic gradient should be maintained to regulate saline intrusion, and a portion
202 of the freshwater should be allowed to flow into the sea. This hydraulic gradient to the seaward
203 acts as a hydraulic barrier towards the SWI (Todd, D. K. & Mays, 2005; Tsanis & Song, 2001;
204 van Dam, 1999). Different methodologies to regulate SWI have been attempted by (Bruington,
205 1972; Pool & Carrera, 2010; Richter, 1953; Sharma & Sanghi, 2013; M. M. Sherif & Hamza,
206 2001; Todd, 2015; van Dam, 1999). Seawater intrusion can be controlled by several methods
207 using biological, chemical, and physical techniques. It is an expensive process and can take a
208 long time, depending on the source of intrusion and level of salinization. Different measures
209 have been presented to control SWI in coastal aquifers. (Todd, 1980) presented various
210 methods of preventing saltwater from contaminating groundwater sources, and later it was
211 revised by (Todd, D. K. & Mays, 2005) to control seawater intrusion.

212 The method used to control saltwater intrusion in coastal aquifers:

- 213 • Changing the pumping pattern (reduction of pumping rates and relocation of pumping
214 wells)
- 215 • Use of physical surface/subsurface barriers,
- 216 • Natural/ artificial recharge (pressure or positive barriers),
- 217 • Pumping of saline water along with the seacoast (abstraction or negative barriers) a
- 218 • Combination techniques (mixed barriers).

219 With high rates of saline water pumping, some of these methods require the disposal of
220 abstracted salts to reduce environmental problems (Bruington, 1972).

221 Each method has advantages and weaknesses when compared to the others. One method might
222 be suitable for one specific location, while another location might not. With the several methods
223 mentioned above to control the seawater intrusion, the economy plays a significant role.
224 Reducing pumping and rearranging pumping wells are cost-effective but not practicable. The
225 abstraction of seawater and a subsurface barrier are expensive methods. Hence they are
226 generally not considered. Increasing groundwater recharge by percolation ponds is
227 economically feasible and quickly adopted(Purnama & Marfai, 2012).

228 **5.1 REDUCTION OF PUMPING RATES**

229 Reduced abstraction from pumping wells could be the most straightforward and cost-effective
230 way to maintain aquifer groundwater balance and monitor SWI. However, the options for
231 reducing abstraction may be limited in certain areas due to water demand requirements. A
232 backup source of water must be given to compensate for the forced reduction in the pumping
233 schedule (Bruington, 1972). Increasing public understanding of the importance of minimizing
234 water losses in water use and supply network systems in various residential, agricultural, and
235 industrial sectors, as well as encouraging them to use renewable or recycling services (e.g.,
236 recycled treated wastewater and desalinated water), will significantly contribute to the
237 scenario's success. However, the cost of providing good quality water from these resources and

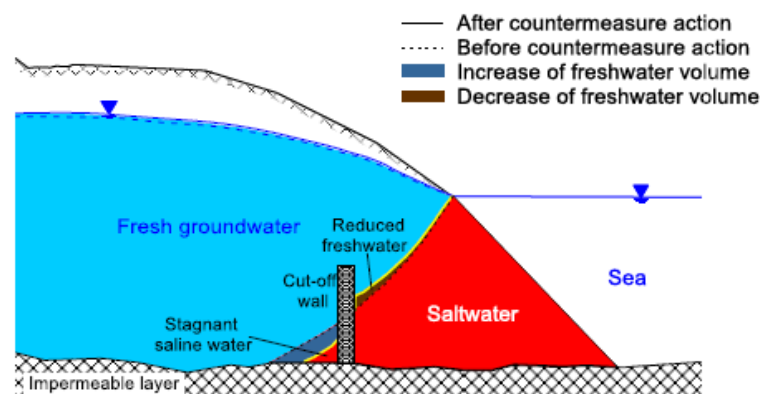
238 making it accessible in vulnerable areas is another problem that limits this approach to a
 239 temporary solution for aquifer protection. This is especially true in areas experiencing rapid
 240 population growth and, as a result, rising demand patterns that are unable to achieve the desired
 241 results (Abd-Elhamid, H. & Javadi, 2008; Vandenbohede et al., 2009) suggested that the actual
 242 conditions of the SWI problem be regulated by integrating this temporary management strategy
 243 with other control steps. (Reichard & Johnson, 2005) demonstrated that using this approach
 244 could help reduce the amount (and cost) of in-lieu delivery of surface water for general
 245 consumption by residents and feed the aquifer against SWI. (Bear, 1999) demonstrated that
 246 redistribution of pumping plans, such as minimizing or turning off wellfields that are less
 247 vulnerable to intrusion, can help reduce SWI in the Nile delta aquifer.

248 5.2 RELOCATION OF PUMPING WELLS

249 Pumping wells are usually moved further inland away from the coast in this method to provide
 250 a stronger seaward hydraulic gradient by holding groundwater levels above sea level and
 251 reducing the excessive outflow of fresh groundwater (Todd, D. K. & Mays, 2005; van Dam,
 252 1999). This approach may be restricted in some cases due to land scarcity or conflicts with other
 253 public-sector or private-sector strategic initiatives, both of which would put an end to the
 254 process. The size (length) of the aquifer to accommodate new positions of wells far from the
 255 intruded saltwater wedge is another barrier to managing SWI in some cases and under high
 256 levels of pollution. Furthermore, the expense of transporting and in-lieu delivering water from
 257 new pumping wells (far enough from the coast) to pre-developed areas near the seashore
 258 (susceptible zones) may be a constraint in this strategy. As a result, this technique is also a
 259 temporary solution and does not prevent saline water intrusion into the aquifer (Abd-Elhamid,
 260 2010). (Sherif, M. M. & Al-Rashed, 2001) suggested planning pumping in the center of the Nile
 261 delta aquifer (simulated in 2D aerial using SUTRA) to ensure potential sustainability.
 262 (Maimone, M. & Fitzgerald, 2001) emphasized the management strategy for a coastal aquifer
 263 by installing a new well system away from the coastal region as a more reliable technique than
 264 another solution that assumed a more profound drilling plan for the old-existing well fields
 265 along the coast. The system's old wells extracted brackish water, which was then used for
 266 domestic purposes after desalination

267 5.3. USE OF PHYSICAL SURFACE / SUBSURFACE BARRIERS,

268 Concrete, grout, bentonite, slurry walls, and sheet piles are widely used in physical subsurface
 269 barriers in front of seawater along the shore (M. S. Hussain et al., 2019).



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Figure 5 : Physical surface/subsurface barriers(Hussain et al., 2019)

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The physical barrier demonstrated its great defensive ability when placed at greater depths and closer to the shore in the absence of abstraction. Under this pumping action, and in aquifer systems where the inland boundary is defined with a lower inflow velocity than the intruded velocity of saline water and high anisotropy, physical barrier control design is highly beneficial in protecting the aquifer. In general, installing a physical subsurface barrier is costly in deep aquifers (Oude Essink, 2001). Another problem has been the stagnant state of the intruding saltwater that has remained behind the barrier since its construction (Figure 5). However, under heavy pumping and a low hydraulic gradient of the flow system, the reverse diffusion of seawater through the barrier and underneath its foundation, and thus mixing with this plume of stagnant saltwater, may cause a process deficiency. (James et al., 2001) introduced a biological barrier against SWI in a completely different and novel technique. The technique relied on the injection of bacteria and nutrient solutions to lower the hydraulic conductivity of subsurface layers and thus the risk of SWI .

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5.4 ARTIFICIAL RECHARGE

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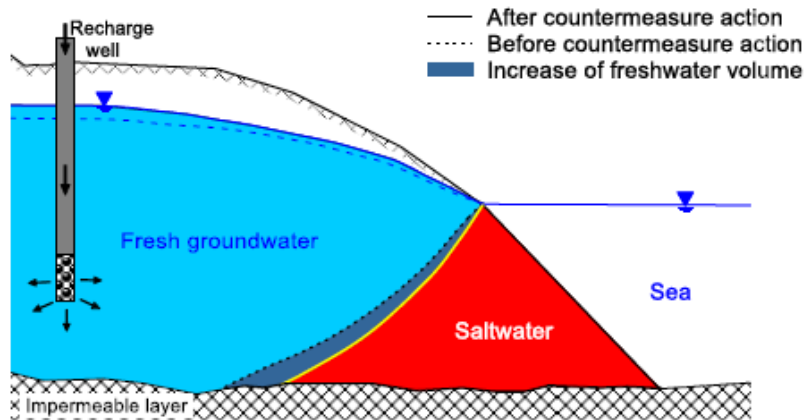
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SWI may occur in any geological state of a confined or unconfined aquifer with a single or multiple layers (Figure). The aquifer is artificially recharged by high-quality water (e.g., surface water, rainwater, pumped groundwater, treated wastewater, or desalinated water) within the positive or pressure barriers to preserve the system's seaward gradient by raising the inland piezometric heads (M. S. Hussain et al., 2019). In general, artificial water recharge aims to reduce flood flows, store water in aquifers, increase groundwater levels, relieve over-pumping, and eventually improve water quality and suppress saltwater (Lahr, 1982). The approach is one of the most recommended and evaluated strategies in the literature. However, in the case of intensive pumping, the use of artificial recharge can be unsuccessful (K. A. Narayan et al., 2007).



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Figure 6: Artificial Recharge (Hussain et al., 2019)

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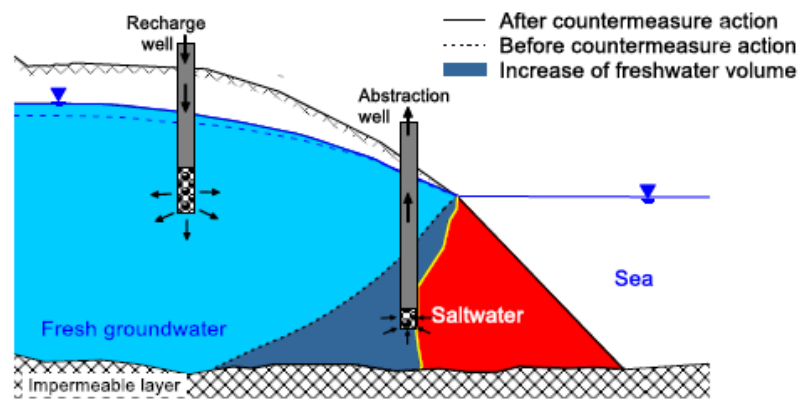
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(Luyun et al., 2011) investigated the positive potential of the deep recharge barrier in SWI repulsion using experimental, computational, and numerical modeling. Surface reservoirs, dams, canals, and other spreading recharge basins can also be used as recharge systems to feed unconfined aquifer systems by collected water infiltration. One of the key constraints to the

302 recharge barriers is the expense of supplying high-quality water (e.g., desalinated water) and its
 303 delivery for recharging purposes. Furthermore, the lack of such water locally, especially during
 304 dry years or in water-stressed areas, maybe a significant constraint (H. F. Abd-Elhamid &
 305 Javadi, 2011). As a result, in recent years, a greater focus has been put on sustainable water
 306 sources, such as treated wastewater, as a means of recharge against seawater intrusion (e.g
 307 (Hussain, 2015; Javadi et al., 2015; Kourakos & Mantoglou, 2011; Koussis et al., 2010;
 308 Reichard & Johnson, 2005; Shammas & Thunvik, 2009). The use of recycled water in common
 309 utility sectors and artificial storage in subsurface layers will help to meet a portion of the water
 310 demand, resist flooding and drought, and protect the system from SWI.

311 5.5 COMBINED BARRIER

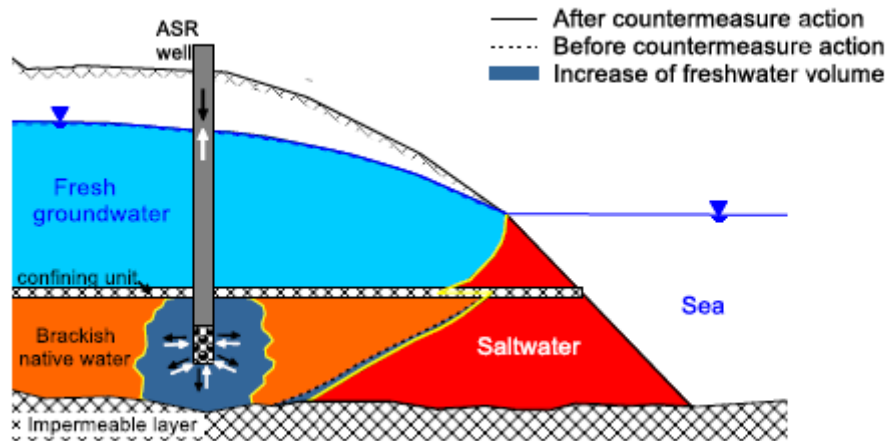
312 Combining both above techniques will help improve control of saltwater intrusion by
 313 combining the advantages of the individual strategies. For example, a combination of pumping
 314 rate reduction and recharge barrier (Johnson et al., 2001) and pumping rate control with
 315 artificial recharge The mixture of positive and negative obstacles is referred to as a mixed or
 316 hybrid barrier (Figure 8). The technique is concerned with the specification of both extraction
 317 and recharge. This mixed barrier (the primary focus of this section) employs the principles of
 318 a) repeated cycles of recharge and abstraction via the same well system and b) injection of
 319 freshwater while abstracting saline water. Several studies have indicated that variants of the
 320 above scheme are the most successful approach among other forms of barriers (Abd-Elhamid,
 321 2010; Kourakos & Mantoglou, 2013; Koussis et al., 2010; Rastogi et al., 2004; S. Troisi, 1994;
 322 Tsanis & Song, 2001)



324 Figure 7 : Combined Barrier (Hussain et al., 2019)

325 5.5.1 AQUIFER STORAGE AND RECOVERY (ASR)

326 Aquifer Storage and Recovery (ASR) is a technique introduced by (Cederstrom, 1947) and
 327 has been widely used in developed countries for the management of water resources as an
 328 alternative to surface storage of water such as dams and reservoirs (Hussain et al., 2019). The
 329 methodology involves persistent storage of excess water by deep injection through recharge
 330 wells into a deep aquifer or other water-bearing formations when water is available or during
 331 the year's wet and low-demand season (Figure 8).

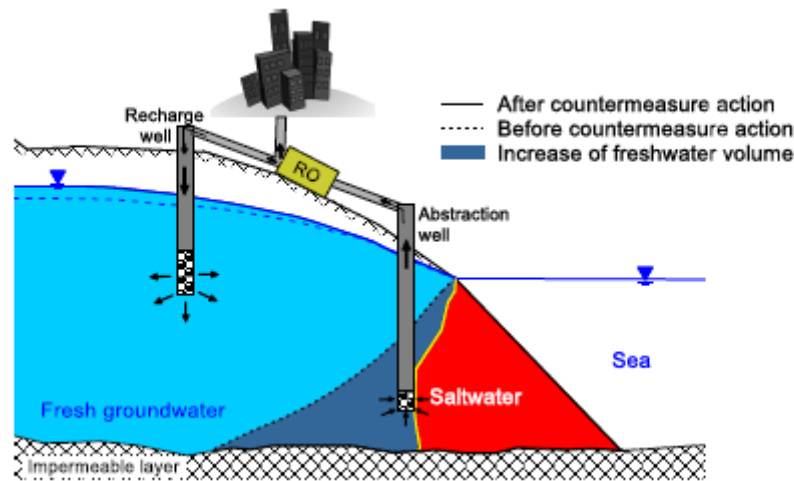


333 Figure 8 : Aquifer Storage and Recovery (Hussain et al., 2019)

334 The stored water is then recovered when needed using the same wells to meet the community's
 335 water demand and during the next dry or high-demand season of the year (Pyne, 1995).
 336 Depending on the quality and level of contamination of the native groundwater in the aquifer
 337 and the recharge water quality, the recovered water may be required to pass a short treatment
 338 process. The ASR process is repeated continuously over the years depending on the levels of
 339 water scarcity. The recovery efficiency in the ASR process is defined as the total volume of
 340 recovered water as a percentage of the volume stored in each operating cycle while satisfying
 341 a target water quality criterion in the abstracted water (Pyne, 1995). Like any other deep
 342 management strategy, ASR has some economic cost and quality limitations, and its
 343 implementation needs further attention. For instance, the high rates of mixing of injected water
 344 with a poor-quality native groundwater tend to reduce the total recoverable volume of
 345 freshwater (recovery efficiency). In addition to this water-water interaction, the quality of the
 346 injected water can also be threatened by water-rock hydrochemical interaction(Hussain, 2015).

347 **5.5.2 ABSTRACTION, DESALINATION AND RECHARGE (ADR)**

348 An innovative and cost-effective methodology called the Abstraction, Desalination, and
 349 Recharge (ADR) is used to control saltwater intrusion in coastal aquifers (H. F. Abd-Elhamid
 350 & Javadi, 2011). This methodology aims to overcome all or at least most of the limitations of
 351 the previous models figure 9.



353 Figure 9 : Abstraction, Desalination, and Recharge (ADR) Methodology(Hussain et al., 2019)

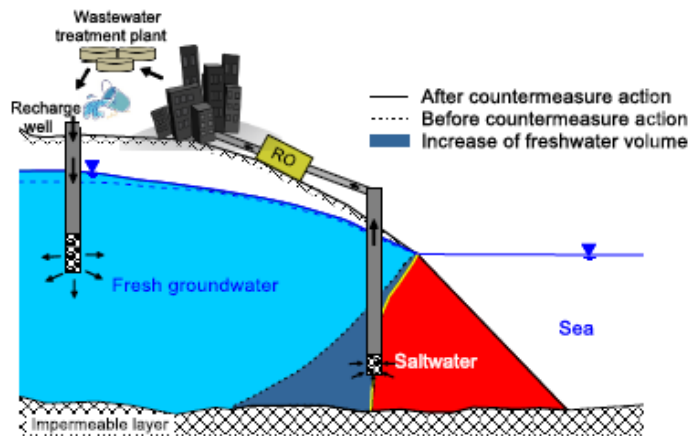
354 Abstraction, Desalination, and Recharge (ADR) consists of three components:

- 355 • Abstraction (A) of brackish water from the saline zone using abstraction wells.
- 356 • Desalination (D) of abstracted brackish water using reverse osmosis process, and
- 357 • Recharge(R) of the treated water into the aquifers using recharge wells.

358 They demonstrated that this approach is both efficient and cost-effective in controlling SWI.
 359 This methodology's significant advantages in controlling SWI have been lower energy
 360 consumption, lower cost, and lower environmental impact. Desalination of the harvested saline
 361 water resulted in an improvement in overall usable freshwater due to the technique. It will be
 362 impossible to use desalinated water to restore the aquifer in arid or semi-arid regions where the
 363 combined amount of desalinated water and abstracted freshwater does not meet demand. On
 364 the other hand, desalinated water is a relatively large water source generated by an energy-
 365 intensive process. The ADR methodology is not effective when the mixing zone/ interface is
 366 close to the sea. Desalination may involve water with very high salinity, which could increase
 367 desalination costs. In this case, the abstraction of saline water and disposal to the sea could be
 368 more cost-effective. Also, if a sufficient freshwater source is available in the area, freshwater
 369 recharge could be more efficient. Low-cost effluents reclaimed from urban wastewater
 370 treatment plants can recharge the aquifer to solve this issue.

371 5.5.3 ABSTRACTION, DESALINATION, AND RECHARGE BY TREATED 372 WASTEWATER (ADRTWW))

373 This methodology uses the more cost-effective water sources such as chemically treated
 374 wastewater (TWW), collected rainwater, and the transfer of good-quality water from nearby
 375 rivers, canals, reservoirs, and ponds. In developed urban areas, treated wastewater (TWW) will
 376 provide a reliable, long-term water source. The availability of other outlets, on the other hand,
 377 is determined by the coastal site's pre-existing hydrological features and hydro-environmental
 378 formations. As a result, TWW is the main target of the proposed process, ADRTWW
 379 (abstraction, desalination, and recharge by TWW), which is shown schematically in Figure 10.



382 Figure 10 : Abstraction, Desalination, and Recharge with treated wastewater (ADRTWW)
 383 Methodology (Hussain et al., 2019)

384 The ADRTWW has a greater potential for monitoring SWI than ADR and traditional barriers
 385 because of its low economic and environmental costs (isolated or combined).

386 Three stages comprise the ADRTWW methodology:

- 387 • Using deep negative barriers (abstraction wells) to remove brackish water from the
 388 saline wedge;
- 389 • Desalination of the collected brackish water using a small-scale reverse osmosis (RO)
 390 plant and use of the desalinated water as a replacement for the urban water supply
 391 system; and
- 392 • Recharging the aquifer using an external source of TWW (Hussain, 2015).

393 The main differences between the ADRTWW and ADR methodologies:

- 394 i. In ADR, the excess desalinated water is used to recharge the aquifer, while in
 395 ADRTWW, the entire desalinated water is used for public use. In ADRTWW, high-
 396 quality recycled water is used for recharge instead of desalinated water, which has a
 397 lower economic cost.
- 398 ii. In the ADR method, the aquifer is recharged by deep injection wells. In the ADRTWW
 399 approach, unconfined aquifers may also be recharged by surface infiltration basins (with
 400 lower energy consumption).

401 6. DISCUSSIONS

402 Seawater intrusion is a natural process that develops over time because of overexploitation of
 403 coastal aquifers and direct hydraulic contact between freshwater and saltwater bodies of water.
 404 Coastal groundwater systems that have not been altered are less susceptible to this occurrence.
 405 Pumping techniques accelerate the process of Seawater Intrusion. Seawater intrusion into
 406 coastal aquifers can be controlled in a variety of ways including pumping control, artificial
 407 recharge, re-location of pumping and well fields, management of landuse, use of scavenger
 408 wells. The critical component in reducing seawater intrusion is maintaining an appropriate
 409 balance of abstraction and recharge in the catchment area and its surroundings. In general, a
 410 seaward hydraulic gradient should be maintained, while allowing a portion of natural freshwater

411 recharge to flow into the sea. The required groundwater flux to the sea can be maintained by
412 properly managing coastal aquifers and limiting the seawater's ultimate landward reach. The
413 optimum discharge of freshwater to the sea can be determined in order to maintain the interface
414 (dispersion zone) in that position. Numerous studies examined the application of pumping
415 control methods (Ahlfeld & Heidari, 1994; Gorelick, 1983; M. Sherif & Kacimov, 2007). While
416 the alternatives outlined above may be practical for preventing seawater intrusion, they may not
417 be economically or environmentally viable due to natural environmental circumstances.
418 Numerous approaches examined in earlier research for preventing and managing seawater
419 intrusion in the study area were excluded from consideration, including the proposed technique.
420 Pumping brackish water is a proposed technique for restoring the balance between freshwater
421 and saline water in coastal aquifers in order to reduce the problem of seawater intrusion.
422 Brackish water can be pumped from the dispersion zone at a specified discharge rate for a
423 specified period of time and used for desalination or to irrigate crops with a higher salt
424 tolerance. This type of abstraction is expected to prevent saline water intrusion, which can result
425 in the dispersion zone being pushed closer to the coast. This process will be repeated until a
426 state of dynamic equilibrium with regard to the salinity distribution is achieved. Pumping
427 brackish groundwater from coastal aquifers would decrease seawater migration deep into the
428 aquifer and contribute to improving groundwater quality (M. M. Sherif & Hamza, 2001).
429 (Rastogi et al., 2004) reported that by combining freshwater injection with saline water
430 extraction, the volume of seawater can be reduced while the volume of freshwater increased.
431 The authors constructed a series of three-dimensional numerical simulations utilizing
432 multidimensional hydrodynamic dispersion models. (Park et al., 2012). They explored a range
433 of seawater extraction systems for mitigating seawater intrusion caused by groundwater
434 pumping in their study. This approach of pumping brackish water to prevent seawater intrusion
435 is not well known at the moment due to a knowledge gap about the needed discharge rates at
436 optimal pumping locations and time duration analyses (M. M. Sherif & Hamza, 2001). To our
437 knowledge, few studies have been undertaken to determine the efficacy of pumping brackish
438 water to minimize seawater intrusion. The studies undertaken thus far have emphasized its
439 potential as a means of mitigating seawater intrusion but have not conducted a complete
440 examination of its effectiveness in light of many criteria such as pumping site, abstraction rates
441 and intensity, length required, and cost analysis. The optimal pumping position required to
442 reduce or shift the saline water toward the shore in order to eventually maintain or push the
443 dispersion zone toward the coast has not been specified in any of these investigations.
444 According to (M. M. Sherif & Hamza, 2001), brackish groundwater pumping can be used to
445 regulate the interface between fresh and subsurface saline water bodies. A well-designed
446 brackish water abstraction system located some distance from the shore would provide a
447 stagnant interface. Thus, as long as the balance between recharge and discharge is maintained,
448 more freshwater can be pumped.(van Dam, 1999) discussed this type of brackish/fresh
449 pumping combination in coastal aquifers. He also commented on the usage of scavenger wells
450 as a mechanism for local and regional scale control of the interface's position and movement.
451 (Sowe et al., 2019) would use pumping brackish water techniques to enhance groundwater
452 quality and may theoretically be used pumping brackish water to neighbouring desalination
453 plants.

455 **7. CONCLUSION AND RECOMMENDATION**

456 This research provided a thorough review of the most common approaches for controlling and
457 mitigating SWI in coastal aquifers. The approaches given include some conventional practice
458 that can be used as temporary solutions before the main control scheme is implemented.
459 Another way of control is to install surface and subsurface physical barriers in front of the
460 intruding seawater body. Hydraulic barriers, on the other hand, have grown in popularity as a
461 result of the high costs of equipment, materials, and large-scale installation of physical barriers.
462 Different combinations of these control strategies can be implemented reduce the risk of
463 seawater intrusion and its impact on the quality of groundwater in the coastal aquifer. The
464 implementation seawater intrusion knowledge, such as processes, investigative methodologies,
465 and managerial approaches, into practice is perhaps the most critical scientific problem in
466 seawater intrusion research. To close the significant gaps between seawater intrusion
467 knowledge and management practice, strategies are required. To illustrate the benefits of
468 seawater intrusion research methodologies and knowledge, more studies that retrospectively
469 and critically analyze the effectiveness of management practices, in particular initiatives
470 deriving from SWI research findings, are needed

471 An environmental impact assessment should be conducted during the planning stage to
472 highlight the environmental impacts of various control measures. The benefits and drawbacks
473 of each methodology are discussed in this study. Controlling SWI in a cost-effective and
474 efficient manner should be a priority for every groundwater exploration project in coastal areas.
475 Furthermore, limitations in terms of economic cost and water demands of increasing population
476 in coastal areas should be considered. While the alternatives outlined above may be practical
477 for preventing seawater intrusion, they may not be economically or environmentally viable due
478 to natural environmental circumstances. Numerous approaches examined in earlier research for
479 preventing and managing seawater intrusion in the study area were excluded from
480 consideration, including the proposed technique. Pumping brackish water is a
481 proposed technique for restoring the balance between freshwater and saline water in coastal
482 aquifers in order to reduce the problem of seawater intrusion research on the most appropriate
483 desalination technology for the location which could improve the water quality and reduce both
484 the economic cost and environmental impact especially on marine environment. Subsurface
485 intakes were not considered a mitigating strategy for seawater intrusion into costal aquifers in
486 the available literature and experiments.

487

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495

496 **8. REFERENCE**

- 497 Abd-Elhamid. (2010). A Simulation-Optimization Model to Study the Control of Seawater
498 Intrusion in Coastal Aquifers. In *Engineering*. University of Exeter.
- 499 Abd-Elhamid, H. & Javadi, A. (2008). *Intrusion in Coastal Aquifers*. 790–797.
500 [https://doi.org/10.1061/40972\(311\)98](https://doi.org/10.1061/40972(311)98)
- 501 Abd-Elhamid, H. F., & Javadi, A. A. (2011). A Cost-Effective Method to Control Seawater
502 Intrusion in Coastal Aquifers. *Water Resources Management*, 25(11), 2755–2780.
503 <https://doi.org/10.1007/s11269-011-9837-7>
- 504 Abd-Elhamid, H., Javadi, A., Abdelaty, I., & Sherif, M. (2016). Simulation of seawater
505 intrusion in the Nile Delta aquifer under the conditions of climate change. *Hydrology*
506 *Research*, 47(6), 1198–1210. <https://doi.org/10.2166/NH.2016.157>
- 507 Ahlfeld, D. P., & Heidari, M. (1994). Applications of Optimal Hydraulic Control to Ground-
508 Water Systems. *Journal of Water Resources Planning and Management*, 120(3), 350–365.
509 [https://doi.org/10.1061/\(asce\)0733-9496\(1994\)120:3\(350\)](https://doi.org/10.1061/(asce)0733-9496(1994)120:3(350))
- 510 Alexander Herzberg. (1901). “Die Wasserversorgung einiger Nordseebäder [The water supply
511 of some North Sea spas]” by Alexander Herzberg (1901). *Hydrogeology Journal*, 26(6),
512 1789–1799. <https://doi.org/10.1007/s10040-018-1772-8>
- 513 Aria, M., & Cuccurullo, C. (2017). bibliometrix: An R-tool for comprehensive science mapping
514 analysis. *Journal of Informetrics*, 11(4), 959–975.
515 <https://doi.org/10.1016/J.JOI.2017.08.007>
- 516 Barlow, P. M. (2003). Ground water in freshwater-saltwater environments of the Atlantic Coast.
517 In *Circular*. <https://doi.org/10.3133/cir1262>
- 518 Barlow, P. M., & Reichard, E. G. (2010). L’intrusion d’eau salée dans les régions côtières
519 d’Amérique du Nord. *Hydrogeology Journal*, 18(1), 247–260.
520 <https://doi.org/10.1007/S10040-009-0514-3>
- 521 Bear. (1999). *Seawater Intrusion in Coastal Aquifers — Concepts, Methods and Practices* (J.
522 Bear, A. H.-D. Cheng, S. Sorek, D. Ouazar, & I. Herrera, Eds.; Vol. 14). Springer
523 Netherlands. <https://doi.org/10.1007/978-94-017-2969-7>
- 524 Bear, J. (1985). Motion of the seawater interface in a coastal aquifer by the method of successive
525 steady states. *Journal of Hydrology*, 76(1–2), 119–132.
- 526 Bear, J. & Zhou, Q. (2007). Sea water intrusion into coastal aquifers. *The Handbook of*
527 *Groundwater Engineering*, Second Edition.
528 <https://doi.org/10.1201/9781420006001.ch12>
- 529 Bear, J. J. & Cheng, H. D. A. (1999). *Conceptual and Mathematical Modeling* (pp. 127–161).
530 https://doi.org/10.1007/978-94-017-2969-7_5
- 531 Bruington, A. E. (1972). SALTWATER INTRUSION INTO AQUIFERS. *Journal of the*
532 *American Water Resources Association*, 8(1), 150–160. [https://doi.org/10.1111/j.1752-](https://doi.org/10.1111/j.1752-1688.1972.tb05104.x)
533 [1688.1972.tb05104.x](https://doi.org/10.1111/j.1752-1688.1972.tb05104.x)

- 534 Cao, T., Han, D., Song, X., & Trolle, D. (2020). Subsurface hydrological processes and
535 groundwater residence time in a coastal alluvium aquifer: Evidence from environmental
536 tracers ($\delta^{18}\text{O}$, $\delta^2\text{H}$, CFCs, ^3H) combined with hydrochemistry. *Science of the Total*
537 *Environment*, 743. <https://doi.org/10.1016/j.scitotenv.2020.140684>
- 538 Cederstrom, D. J. (1947). Artificial recharge of a brackish water well. *The Commonwealth*,
539 *December 1*, 6.
- 540 Drabbe, J., & Ghijben, B. W. (1889). Nota in verband met de voorgenomen putboring nabij.
541 *Tijdschrift van Het Koninklijke Instituut van Ingenieurs*, 8, 22.
- 542 Ferguson, G., & Gleeson, T. (2012). Vulnerability of coastal aquifers to groundwater use and
543 climate change. *Nature Climate Change*, 2(5), 342–345.
544 <https://doi.org/10.1038/NCLIMATE1413>
- 545 Garfield, E. (2009). From the science of science to Scientometrics visualizing the history of
546 science with HistCite software. *Journal of Informetrics*, 3(3), 173–179.
- 547 Gorelick, S. M. (1983). A review of distributed parameter groundwater management modeling
548 methods. *Water Resources Research*, 19(2), 305–319.
549 <https://doi.org/10.1029/WR019i002p00305>
- 550 Han, D., & Currell, M. J. (2018). Delineating multiple salinization processes in a coastal plain
551 aquifer, northern China: Hydrochemical and isotopic evidence. *Hydrology and Earth*
552 *System Sciences*, 22(6), 3473–3491. <https://doi.org/10.5194/hess-22-3473-2018>
- 553 Hassan, S. U., Haddawy, P., & Zhu, J. (2014). A bibliometric study of the world's research
554 activity in sustainable development and its sub-areas using scientific literature.
555 *Scientometrics*, 99(2), 549–579. <https://doi.org/10.1007/S11192-013-1193-3>
- 556 Howard, K. w. F. (1987). Beneficial Aspects of Sea-Water Intrusion. *Ground Water*, 25(4),
557 398–406. <https://doi.org/10.1111/j.1745-6584.1987.tb02144.x>
- 558 Hubbert, M. K. (1940). Theory of Groundwater Motion. *Journal of Geology*, 8, 785–944.
559 <https://doi.org/http://dx.doi.org/10.1086/624930>
- 560 Hussain, M. S. (2015). *Numerical Simulation and Effective Management of Saltwater Intrusion*
561 *in Coastal Aquifers*.
- 562 Hussain, M. S., Abd-Elhamid, H. F., Javadi, A. A., & Sherif, M. M. (2019). Management of
563 Seawater Intrusion in Coastal Aquifers: A Review. *Water*, 11(12), 2467.
564 <https://doi.org/10.3390/W11122467>
- 565 James, G., Hiebert, R., Warwood, B., & Cunningham, A. (2001). *Subsurface Biofilm Barriers*
566 *for Controlling Saltwater Intrusion*. 3–6.
- 567 Javadi, A., Hussain, M., Sherif, M., & Farmani, R. (2015). Multi-objective Optimization of
568 Different Management Scenarios to Control Seawater Intrusion in Coastal Aquifers. *Water*
569 *Resources Management*, 29(6), 1843–1857. <https://doi.org/10.1007/S11269-015-0914-1>
- 570 Johnson, T., Reichard, E., Land, M., & Crawford, S. (2001). Monitoring , Modeling , and
571 Managing Saltwater Intrusion , Central and West Coast Groundwater Basins , Los Angeles

- 572 County , California. *First International Conference on Saltwater Intrusion and Coastal*
573 *Aquifers*, 2–3.
- 574 Kourakos, G., & Mantoglou, A. (2011). Simulation and Multi-Objective Management of
575 Coastal Aquifers in Semi-Arid Regions. *Water Resources Management*, 25(4), 1063–
576 1074. <https://doi.org/10.1007/s11269-010-9677-x>
- 577 Kourakos, G., & Mantoglou, A. (2013). Development of a multi-objective optimization
578 algorithm using surrogate models for coastal aquifer management. *Journal of Hydrology*,
579 479, 13–23. <https://doi.org/10.1016/j.jhydrol.2012.10.050>
- 580 Koussis, A. D., Georgopoulou, E., Kotronarou, A., Mazi, K., Restrepo, P., Destouni, G., Prieto,
581 C., Rodriguez, J. J., Rodriguez-Mirasol, J., Cordero, T., Ioannou, C., Georgiou, A.,
582 Schwartz, J., & Zacharias, I. (2010). Gestion économe des aquifères côtiers via la recharge
583 d’eaux usées traitées et la désalinisation d’eaux souterraines saumâtres: Application au
584 bassin et a l’aquifère d’Akrotiri, Chypre. *Hydrological Sciences Journal*, 55(7), 1234–
585 1245. <https://doi.org/10.1080/02626667.2010.512469>
- 586 Leydesdorff, L., & Bihui, J. (2005). Mapping the chinese science citation database in terms of
587 aggregated journal-journal citation relations. *Journal of the American Society for*
588 *Information Science and Technology*, 56(14), 1469–1479.
589 <https://doi.org/10.1002/ASI.20209>
- 590 Luyun, R., Momii, K., & Nakagawa, K. (2011). Effects of Recharge Wells and Flow Barriers
591 on Seawater Intrusion. *Ground Water*, 49(2), 239–249. <https://doi.org/10.1111/j.1745-6584.2010.00719.x>
- 593 Mabrouk, M., Jonoski, A., Oude Essink, G. H. P., & Uhlenbrook, S. (2018). Impacts of sea
594 level rise and groundwater extraction scenarios on fresh groundwater resources in the Nile
595 Delta Governorates, Egypt. *Water (Switzerland)*, 10(11).
596 <https://doi.org/10.3390/w10111690>
- 597 Maimone, M. & Fitzgerald, R. (2001). Effective Modeling of Coastal Aquifer Systems. *First*
598 *International Conference on Saltwater Intrusion and Coastal Aquifers- Monitoring,*
599 *Modeling, and Management. Essaouira-Morocco.*
- 600 Niu, B., Loáiciga, H. A., Wang, Z., Zhan, F. B., & Hong, S. (2014). Twenty years of global
601 groundwater research: A Science Citation Index Expanded-based bibliometric survey
602 (1993-2012). *Journal of Hydrology*, 519(PA), 966–975.
- 603 Oude Essink, G. H. P. (2001). Improving fresh groundwater supply - Problems and solutions.
604 *Ocean and Coastal Management*, 44(5–6), 429–449. [https://doi.org/10.1016/S0964-5691\(01\)00057-6](https://doi.org/10.1016/S0964-5691(01)00057-6)
- 606 Park, Y., Lee, J. Y., Kim, J. H., & Song, S. H. (2012). National scale evaluation of groundwater
607 chemistry in Korea coastal aquifers: Evidences of seawater intrusion. *Environmental Earth*
608 *Sciences*, 66(3), 707–718. <https://doi.org/10.1007/S12665-011-1278-3>

- 609 Pool, M., & Carrera, J. (2010). Dynamics of negative hydraulic barriers to prevent seawater
610 intrusion. *Hydrogeology Journal*, 18(1), 95–105. [https://doi.org/10.1007/s10040-009-](https://doi.org/10.1007/s10040-009-0516-1)
611 0516-1
- 612 Post, V. E. A. (2005). Fresh and saline groundwater interaction in coastal aquifers: Is our
613 technology ready for the problems ahead? *Hydrogeology Journal*, 13(1), 120–123.
614 <https://doi.org/10.1007/S10040-004-0417-2>
- 615 Post, V. E. A. (2018). Annotated translation of “Nota in verband met de voorgenomen putboring
616 nabij Amsterdam [Note concerning the intended well drilling near Amsterdam]” by J.
617 Drabbe and W. Badon Ghijben (1889). *Hydrogeology Journal*, 26(6), 1771–1788.
618 <https://doi.org/10.1007/s10040-018-1797-z>
- 619 Purnama, S., & Marfai, M. A. (2012). Saline water intrusion toward groundwater : Issues and
620 its control. *Jornal of Natural Resources and Development*, 2, 25–32.
- 621 Pyne, R. D. G. (1995). *Groundwater Recharge and Wells A Guide to Aquifer Storage Recovery*.
622 <https://doi.org/10.1201/9780203719718>
- 623 Qin, L. (2015). *The Bibliometrics Analysis of Related Studies of Urban Development and*
624 *Ecological Safety Condition(2011- 2013)*.
- 625 Rastogi, A., Choi, G. W., & Ukarande, S. K. (2004). Diffused interface model to prevent ingress
626 of seawater in multi-layer coastal aquifers. *J. Special Hydrol.*, 4, 1–31.
- 627 Reichard, E., & Johnson, T. (2005). Assessment of Regional Management Strategies for
628 Controlling Seawater Intrusion. *Journal of Water Resources Planning and Management-*
629 *Asce - J WATER RESOUR PLAN MAN-ASCE*, 131. [https://doi.org/10.1061/\(ASCE\)0733-](https://doi.org/10.1061/(ASCE)0733-9496(2005)131:4(280))
630 9496(2005)131:4(280)
- 631 Richter, B. (1953). Sea-water intrusion into ground-water basins bordering the California coast
632 and inland bays. *Eos, Transactions American Geophysical Union*, 34(4), 575–582.
633 <https://doi.org/https://doi.org/10.1029/TR034i004p00575>
- 634 S. Troisi, R. C. and S. S. (1994). *Sea water intrusion in the coastal aquifer of Reggio Calabria:*
635 *guidelines for management*.
- 636 Seenipandi, K., Nainarpandian, C., Kandathil, R. K., & Sellamuthu, S. (2019). Seawater
637 intrusion vulnerability in the coastal aquifers of southern India—an appraisal of the
638 GALDIT model, parameters’ sensitivity, and hydrochemical indicators. *Environmental*
639 *Science and Pollution Research*, 26(10), 9755–9784. [https://doi.org/10.1007/s11356-019-](https://doi.org/10.1007/s11356-019-04401-0)
640 04401-0
- 641 Shammam, M. I., & Thunvik, R. (2009). Predictive simulation of flow and solute transport for
642 managing the Salalah coastal aquifer, Oman. *Water Resources Management*, 23(14),
643 2941–2963. <https://doi.org/10.1007/s11269-009-9417-2>
- 644 Sharma, S. K., & Sanghi, R. (2013). Wastewater reuse and management. *Wastewater Reuse*
645 *and Management*, 1–500. <https://doi.org/10.1007/978-94-007-4942-9>

- 646 Sherif, M., & Kacimov, A. (2007). *Seawater intrusion in the coastal aquifer of Wadi Ham, UAE*
647 (Vol. 312). IAHS Publ.
- 648 Sherif, M. M. & Al-Rashed, M. F. (2001). Vertical and horizontal simulation of seawater
649 intrusion in the Nile Delta aquifer. *First International Conference on Saltwater Intrusion*
650 *and Coastal Aquifers-Monitoring, Modeling, and Management. Essaouira-Morocco.*
- 651 Sherif, M. M., & Hamza, K. I. (2001). Mitigation of seawater intrusion by pumping brackish
652 water. *Transport in Porous Media*, 43(1), 29–44.
653 <https://doi.org/10.1023/A:1010601208708>
- 654 Sherif, M., & Singh, V. P. (2002). Effect of Groundwater Pumping on Seawater Intrusion in
655 Coastal Aquifers. *Journal of Agricultural and Marine Sciences [JAMS]*, 7(2), 61.
656 <https://doi.org/10.24200/jams.vol7iss2pp61-67>
- 657 Sowe, M. A., Sadhasivam, S., Mostafa Mohamed, M., & Mohsen, S. (2019). Modeling the
658 mitigation of seawater intrusion by pumping of brackish water from the coastal aquifer of
659 Wadi Ham, UAE. *Sustainable Water Resources Management*, 5(4), 1435–1451.
660 <https://doi.org/10.1007/s40899-018-0271-3>
- 661 Todd. (1980). *Groundwater Hydrology-Second Edition.*
- 662 Todd, D. K. (2015). *Salt-Water Intrusion and Its Control Author (s): David K. Todd Source :*
663 *Journal (American Water Works Association), Vol . 66 , No . 3 , Ground Water (March*
664 *Stable URL : <http://www.jstor.org/stable/41266996> . 66(3), 180–187.*
- 665 Todd, D. K. & Mays, L. W. (2005). *Groundwater hydrology* (3rd Editio). Hoboken: John Wiley
666 & Sons.
- 667 Todorovic, A., & Verruijt, E. (1968). A note on the ghyben-herzberg formula. *International*
668 *Association of Scientific Hydrology. Bulletin*, 13(4), 43–46.
669 <https://doi.org/10.1080/02626666809493624>
- 670 Tsanis, I. K., & Song, L.-F. (2001). Remediation of Sea Water Intrusion: A Case Study.
671 *Groundwater Monitoring & Remediation*, 21(3), 152–161.
672 <https://doi.org/https://doi.org/10.1111/j.1745-6592.2001.tb00752.x>
- 673 van Dam, J. C. (1999). *Exploitation, Restoration and Management* (pp. 73–125).
674 https://doi.org/10.1007/978-94-017-2969-7_4
- 675 Vandenbohede, A., van Houtte, E., & Lebbe, L. (2009). Sustainable groundwater extraction in
676 coastal areas: A Belgian example. *Environmental Geology*, 57(4), 735–747.
677 <https://doi.org/10.1007/s00254-008-1351-8>
- 678 Vu, D. T., Yamada, T., & Ishidaira, H. (2018). Assessing the impact of sea level rise due to
679 climate change on seawater intrusion in Mekong Delta, Vietnam. *Water Science and*
680 *Technology*, 77(6), 1632–1639. <https://doi.org/10.2166/WST.2018.038>
- 681 Wei, Z., Lee, X., Aemisegger, F., Benetti, M., Berkelhammer, M., Casado, M., Caylor, K.,
682 Christner, E., Dyroff, C., García, O., González, Y., Griffis, T., Kurita, N., Liang, J., Liang,
683 M.-C., Lin, G., Noone, D., Gribanov, K., Munksgaard, N. C., ... Yoshimura, K. (2019).

684 Data Descriptor: A global database of water vapor isotopes measured with high temporal
685 resolution infrared laser spectroscopy. *Nature Publishing Group*.
686 <https://doi.org/10.1038/sdata.2018.302>

687 Werner, A. D., Bakker, M., Post, V. E. A., Vandenbohede, A., Lu, C., Ataie-Ashtiani, B.,
688 Simmons, C. T., & Barry, D. A. (2013). Seawater intrusion processes, investigation and
689 management: Recent advances and future challenges. *Advances in Water Resources*, *51*,
690 3–26.

691